J. Phycol. 35, 1430-1437 (1999)



MIXOTROPHY AND NITROGEN UPTAKE BY PFIESTERIA PISCICIDA (DINOPHYCEAE)

Alan J. Lewitus, 2 Bonnie M. Willis, Kenneth C. Hayes

Belle W. Baruch Institute for Coastal Research, Baruch Marine Laboratory, University of South Carolina, P.O. Box 1630, Georgetown, South Carolina 29442

JoAnn M. Burkholder, Howard B. Glasgow, Jr.

Department of Botany, Box 7612, North Carolina State University. Raleigh, North Carolina 27695-7612

Patricia M. Glibert

Horn Point Laboratory, University of Maryland Center for Environmental Science, Cambridge, Maryland 21613

and

Marianne K. Burke

USDA Forest Service, Southern Research Station, Charleston, South Carolina 29414

The nutritional versatility of dinoflagellates is a complicating factor in identifying potential links between nutrient enrichment and the proliferation of harmful algal blooms. For example, although dinoflagellates associated with harmful algal blooms (e.g. red tides) are generally considered to be photo. trophic and we inorganic nutrients such as nitrate or phosphate, many of these species also have pronounced heterotrophic capabilities either as osmotrophs or phagotrophs. Recently, the wide pread occurrence of the heterotrophic toxic dinoflage llate, Pfiesteria piscicida Steidinger et Burkholder, has been documented in turbid estuarine waters. Pfiesteria piscicida has a relatively proficient grazing ability, but also has an ability to function as a phototroph by acquiring chloroplasts from algal prey, a process termed kleptoplastidy. We tested the ability of kleptoplastidic P. piscicida to take up 15N-labeled NH;, NO;, urea, or gutamate. The photosynthetic activity of these cultures was verified, in part, by use of the fluorochrome, primulin, which indicated a positive relationship between photosynthetic starch production and growth irradiance. All four N substrates were taken up by P. piscicida, and the highest uptake rates were in the range cited for phytoplankton and were similar to N uptake estimates for phagotrophic P. piscicida. The demonstration of direct nutrient acquisition by kleptoplastidic P. piscicida suggests that the response of the dinoflagellate to nutrient enrichment is complex, and that the specifid pathway of nutrient stimulation(e.g. indirectstimulation through enhancement of phytoplankton prey abundance vs. direct stimulation by saprotrophic nutrient uptake) may depend on P. pisccida's nutritional state (phagotrophy vs. phototrophy).

Key index words: cryptophytes; fish kills; harmful al-

gal bloom; kleptoplastidy; nutrient loading; Pfiesteria piscicida; primulin; starch; toxic dinoflagellates

Pfiesteria piscicida Steidinger et Burkhotder (Dinophyceae) is one of several myzocytotically feeding "heterotrophic" dinoflagellates with a reported capability for kleptoplastidy, the process by which functional chloroplasts are retained from algal prey (Larsen 1992, Laval-Peuto 1992, Schnepf and Elbrächter 1992, Lewitus et al. 1999). Although this form of mixotrophy has been acknowledged in dinoflagellates for a number of years, studies examining the role and regulation of phototrophic and phagotrophic nutrition in kleptoplastidic species are rue (e.g. Fields and Rhodes 1991, Skovgaard 1998). at least partly a consequence of the difficulty in recognizing and culturing kleptoplastidic dinoflagellates (Schnepf et al. 1989, Schnepf and Elbrächter 1992, Lewitus et al. 1999). Schnepf and Elbrächter (1992) and Skovgaard (1998) suggested that dinoflagellates with an inconsistent chloroplast number per cell may be kleptoplastidic, and further speculated that the ability is more common in dinofla**gellates** than previously considered. It is also possible that problems associated with maintaining seemingly "phototrophic" dinoflagellates in a laboratory culture may be related, in some cases, to the unrecognized need to replenish prey supply. Awareness of the prevalence of these and other mixotrophic **protists** has grown dramatically in recent years (Stoecker 1998), stressing the need to understand their physiological ecology and role in microbial food web function.

Stoecker's (1998) classification scheme for mixotrophic protists presented a conceptual model for kleptoplastidic "protozoa" (e.g. some ciliates, sarcodines, and dinoflagellates) based on their relative dependency on phototrophic vs. phagotrophic nutrition, and the response of these processes to light,

Received '7 November 1998. Accepted 4 September 1999.

Author for reprint requests; e-mail Lewitus@belle.baruch.
sc.edu.

dissolved inorganic (nutrients, and particulate food availability. In this model: (a) carbon, nitrogen, and phosphorus are acquired primarily through phagotrophy; (b) growth ivery slow or ceases in the prolonged absence of prey; and (c) photosynthesis is thought to be used to supplement carbon nutrition by covering respiratory demands during periods when prey arc scarce. The data to support the model are based primarily on information from studies with sarcodines or ciliates. Recently, however, Skovgaard (1998) suggested that photosynthesis in kleptoplastidic Gymnodinium 'gracilentum' is used primarily to enhance survival under food limitation, based on the short (-2 day) turnover time of kleptochlo-

Kleptoplastidy also has been demonstrated in Phesteria piscicida (Lewitus et al. 1999), an ichthyotoxic dinoflagellate that, over the last decade, has been implicated as a causative factor of fish kills in North Carolina estuaries and the Chesapeake Bay (Burkholder et al. 1992, 1995, Burkholder and Glasgow 1997, Maryland Department of Natural Resources 1997, University of Maryland Center for Environmental Science 1997). Unlike G. gracilentum, a longer kleptochloroplast retention time (at least 9 days) was observed in P. piscicida, but Lewitus et al. (1999) considered this and the low photosynthetic rates (0.54 pg C-cell-1-h-1) underestimates of photosynthetic capacity because of the limiting growth irradiance used in their experiment. Even at that low irradiance, however, a population doubling was measured, suggesting that phagotrophy was not essential for cell division. The growth, or even maintenance, of photosynthetically active P. piscicida populations implies that nutrients were taken up sapfourophically in support of photoautotrophy. The potential ability of P. piscicida to acquire nutrients directly has implications toward the hypothesized role of nutrient loading in promoting the dinoflagellate's growth and toxic activity (Burkholder et al. 1995, 1997, 1998, Burkholder and Glasgow 1997, Lewitus et al. 1999). That is, recognition that P. pisticida can function as a phytoplankter, and not just as a heterotrophic protist, stresses consideration of nutrient stimulatory mechanisms analogous to those that affect phytoplankton. In this study, we tested the hypothesis that kleptoplastidic P. piscicida can take up N nutrients directly.

MATERIALS AND METHODS

The Phesteria viscicida isolate was obtained from the Neuse River estuary, North Carolina, and its identity was confirmed by thecal plate configuration (Steidinger et al. 1996), using scanning election microscopy (Glasgow, Jr., unpubl.). The nontoxic zoospores used in this experiment were derived from toxic zoospore cultures maintained in fish abuariums. Before beginning the experiment, toxic zoospores were transferred to media (f/2-enriche d filtered seawater; Guillard 1975 but without Si) containing Rhodomonas sp. Karsten CCMP757 (Cryptophyceae), an 8-um diameter cryptophyte previously shown to be a preferred prey species and source of kleptochloroplasts for P. pisaida (Glasgow et al. 1998, Lewitus et al. 1999), Thii culture was maintained at 23° C,

15 ppt, and a light dark cycle of 12:12 h (80 μE·m⁻²·s⁻¹), and inspected microscopically several times a day for P. piscicida feed-

After 3 days of maintenance, P. piscicida was observed to shift from a lag phase, when grazing was not apparent to an active feeding phase when the majority of the population was observed to rapidly ingest cryptophytes (referred to as a "swarming response" or "feeding event"; Glasgow et al. 1998. Lewitus et al. 1999). This feeding event was followed by a reduction in grazing activity, decreased cryptophyte abundance, and increased *P. piscioda* abundance. Five hours after the feeding event, aliquots (73) mL) from this culture were gently transferred to 18 125-ml flasks. Six of these were sampled immediately (designated "initial" treatment), and 12 were placed in the dark. After being kept for 15 h in the dark, the 12 flasks were incubated at a relatively low irradiance (70 µE·m⁻²·s⁻¹) for 3 b, after which six flask, were sampled (designated "low light" treatment). and the remaining sampled (designated light light irradiance (360 µE·m⁻²·s⁻¹) and sampled 3 h later (designated "high light" treatment). The principle of this experimental design was to allow comparison of starch content and N uptake in: (a) "initial" J? piscicida populations (containing recently ingested chloroplasts); (b) "low light" populations (kleptoplastidic cultures incubated for a short rime under a relatively low irradiance); and (c) "high light" populations (kleptoplastidic cultures examined after a shift-up in growth irradiance). These three populations were considered to be klep-toplastidic based on microscopic insspections indicating that grazing on cryptophytes did not occur after the initial feeding event, and examination of DAPI-stained samples indicating that plastid-containing P pisacida did not contain cryptophyte nuclei (Lewitus et al. 1999).

The above cultures were sampled for **P. piscicida** and **Rhodomonas** cell **cound**, particulate **nitrogen**, dissolved nutrients (NO). NO3, dissolved free amino acids (DFAA), urea, and NH:), uptake of ¹³N-labeled NH:, NO3, urea, or olutamate, and dinoflagellate and cryptophyte cell, chloroplast, and starch areas. Cell counts were measured on acid-Lugol's-fixed samples, using a Nageotte Bright Line hemacytometer (depth 0.5 mm). For particulate nitrogen, water was filtered onto precombusted Whatman GF/C filters, frozen, and later measured using a Control Equipment CHN analyzer (Parsons et al. 1984). Disso ved nutrients were measured in the filtrate from sterile 0.45-um polycarbonate membrane filters. NO₅ and NO₅ concentrations were determined colorimetrically using a Technicon autoanalyzer, and urea was measured by the urease method (Parsons et al. 1984). NH2 concentration was determined by the phenol/hypochlorite technique (Solarzano 1969), and DFAA determined by reverse-phase high-performance liquid chromatography (Lindroth and Mopper 1970). Microbar untake material determined using 1910 tracer 1979). Nutrient uptake rates were determined using 13N tracer techniques, following Glibert and Capone (1993). Labeled substrates were added at an initial concentration of 2 µg at N L-1, and samples were incubated for 30 min. Isotopic analyses were conducted using a **Finnigan** MAT 251 mass spectrometer, coupled with a Europa **ANCA** sample inlet system.

Cell morphological parameters were determined using image analysis (Optronics image analysis system with Flashpoint software) on 4',6-diamidino-2-phenylindole (DAPI)- or primulin-stained samples. SampLea were fixed with 0.5% NiSO, (Sieracki et al. 1987) and then 2% hexamethylenetetramine-buffered formaldehyde (Throndsen 1978. Smecker et al. 1989), incubated, with a fluorochrome (DAPI or primulin), filtered, and the filters rapidly frozen on a liquid nitrogen-cooled steel Mock (Lewitus et al. 1999). The **DAPI-staining** method followed Porter and Feig (1980) and was used to determine the percentage of plastid-containing zoospores that also contained ingested cryptophyte nuclei (stained by DAPI). As mentioned DAPI fluorescence indicative of the presence of cryptophyte nuclei inside zoospores was not observed, and therefore these data are not presented. The primulin method followed Caron (1983) Like DAPI, primulin has an excitation maximum in the ultraviolet, but, whereas DAPI stains DNA, primulih reacts with plasmalemma, thecal plates, and starch (Sterling 1964. Revilla et al. 1986, Klut et al. 1989). We used primulin fluorescence to estimate the starch area within 1432

ALAN J. LEWITUS ET AL.

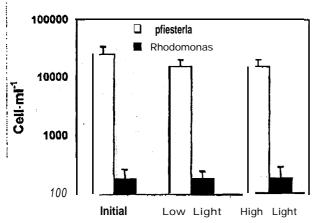


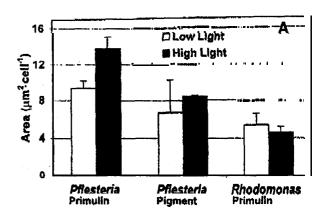
Fig. 1. The mean and standard deviation of *Pfiesteria piscicida* vhite) or *Rhodomonas* sp. (black) abundance in the initial, low sht, or high light cultures.

free-living cryptophyte chloroplasts and zoospore kleptochloroplasts. Starch is a p hotosynthetic product of cryptophytes that is produced within the periplastical space (Santore 1985), and a previous study demonstrated the accumulation of starch within P. piscicida kleptochloroplasts (Lewitus et al. 1999).

In another experiment, we tested the relationship between the primulin-fluorescent area within Cryptomonas sp. HP9001 (Cryptophyceae) cells and the cellular nonstructural carbohydrate in batch cultures grown at three irradiances (211, 104, or 8 µE·m-².s-¹) and sampled from exponential growth phase. Samples were harvested by centrifugation, resuspended in a phosphate buffer, pH 6.8 (Lewitus and Caron 1990), and cell material extracted by cell disruption, using a Biospecs Products Miniberaturated by cell disruption and carbohydrates, using 30% perchloration and cell material by cell disruption and cell material by cell disruption and carbohydrates, using 30% perchloration and cell material by cell disruption and cell material by cell disr



Fig. 2. Epifluorescence micrograph of primulin-stained Pfiesteria pisciaida 2005pore excited by ultraviolet light, showing the kleptochloroplast inside an epithecal vacuole (left). The pigmented area of the kleptochloroplast is reddish orange (phycoerythrin autofluorescence), and the primulin stain shows up as iyellowish white fluorescence surrounding the pigmented area. (scale bar, 4 pm).



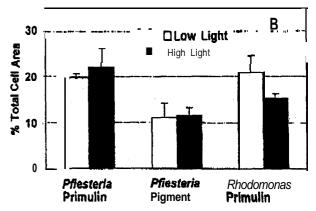


Fig. 3. The mean and standard deviation of (A) area per cell or (B) % total cell area of *Pfusteria piscicida* primulin fluorescence, *Pfusteria piscicida* pigment fluorescence, or *Rhodomonas* sp. primulin fluorescence of low light (white) or high light (black) cultures.

it acid extraction and a colorimetric assay based on the phenolsulfuric acid reaction (Burke et al. 1992). Absorbance was converted to concentration in milligram glucose equivalents Per milliliter using a standard curve (r = 0.9932).

A stest (significance level of 0.05) was used in all comparisons of means referred to below.

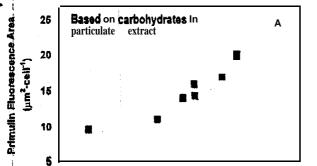
RESULTS

Mean **P. piscicida** population abundance in "initial" cultures (those **sampled 5** h after the **feeding** event) was **nearly 100-fold** greater than **that** of **Rhodomonas** (Fig. 1). **Cryptophyte** abundance did **not** change significantly during the experiment. remaining below 200 **cell·mL-1**. Cultures incubated under "low" or "high" light contained ca. 40% lower **200**-spore abundances than initial cultures, but cell numbers did not **vary** with light treatment.

Primulin-stained samples were analyzed for three variables (Figs. 2, 3): (1) "Pflesteria primulin," the area within the plastid-containing zoospore vacuole characterized by primulin fluorescence (yellow to white area of epitheca in Fig. 2), (2) "Pflesteria pigment," the area within the plastid-containing zoospore vacuole characterized by pigment (phycoerythrin) autofluorescence (reddish orange area in Fig.

0

60



10

20 30 40
Carbohydrate content
(pg glucose equiv *cell")

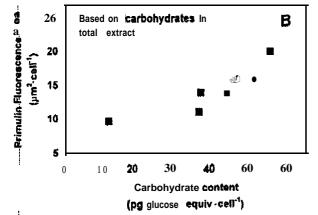
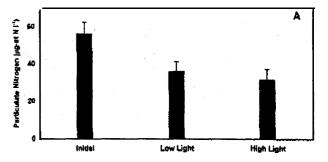


Fig. 4. The relationship between the primulin fluorescence area per cell and nonstructural carbohydrate content per cell in Gryptomonas sp. HP9001 cultures. After total nonstructural carbohydrate extraction, carbohydrate concentrations were measured in both the dissolved (supernatant) and particulate (pellet) fractions. The carbohydrate content is plotted, based strictly on the particulate extract (A) or based on the sum of particulate and dissolved extracts (B).

2), and (3) "Rhodomonas primulin," the area within the free-living **cryptophyte** chloroplast characterized by primulin fluorescence (not included in Fig. 2). In comparing light treatments, the "Pfiesteria primulin" area per zopspore was nearly 50% greater in high-h ght than in low-light cultures, while the amount of "Pfiesteria pigment" or "Rhodomonas primulin" area did not vary with treatment (Fig. 3A). The treatment effect on *Pfiesteria* primulin fluorescence and not *Rhodomonas* primulin fluorescence is evidence that the positive relationship between growth irradiance and primulin-reactive material (presumably starch) is not a function of prey ingestion. Although the primulin fluorescent area per zoospore was significantly greater in cultures grown at high light, the relative proportion of this area to total cell area did not vary significantly with growth irradiance (Fig. 3B). The average area of high- and low-light cells was 54 and 47 μm², respectively (data not shown). The relatively greater absolute, but not



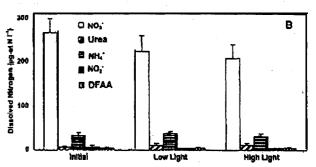


FIG. 5. Mean and standard deviations of (A) particulate nitrogen concentrations and (B) dissolved nitrogen concentrations of initial, low light, or high light cultures. DFAA indicates diilyed free amino acids.

proportional, *Pfiesteria* primulin area per cell in high vs. low light cultures suggests that the increase in cell size was related to an increase in primulin-stained cell material. Our assumption that this material was starch is supported by the strong correlation between primulin fluorescence area and nonstructural carbohydratc content in cells of the cryptophyte. Cryptomonas sp. (Pig. 4). The correlation coefficient between the primulin-stained area per cell and the carbohydrate content was 0.92 or 0.87 when carbohydrates were measured in the pelleted or total extract, respectively.

Particulate nitrogen decreased after incubation at low or high light by an amount (64% or 55%, respectively; Fig. 5A) that corresponded roughly to the decrease in zoospore cell abundance (61% or 59%, respectively; Fig. 1). The dissolved nitrogen pool. was predominantly composed of NO₅ throughout the experiment (Fig. 5B), but the mean ratio of NH₂, urea, and DFAA to NO; concentration increased over time by 41%, 164%, and 62%, respectively, at low light, and 40%, 183%, and 90%, respectively, at high light. These patterns suggest the net biological removal of NO₅ and regeneration of NH₂, urea, and DFAAs over the course of the experiment.

Uptake of all four ¹⁵N-labeled substrates was detected (Table 1). Not surprisingly, given the relatively high NO₅ concentrations in the dissolved N pool, NO; was taken upat the greatest rate, followed by glutamate, NH₅, and urea. Ceils grown at high light took up NH₅ or urea at significantly greater rates than low-light-grown cells, but NO₅ or glu-

ALAN J. LEWITUS ET AL.

TABLE 1. 15N uptake rates (mean ± standard deviaton) of cultures grown at low or high irradiance, expressed as absolute label uptake or uptake normalized to the cell abundance of *Pfiesteria piscicida*.

/	Absolute uptake rate (µg at N·I-1·h-1)				Uptake per Pfinuria pissicida cell (fmole N rell-1 h-1)			
Treament	NO;	ин ;	Urea	Clutamate	NO;	NH;	Urea	Glutamate
Low light High light	2.9 ± 0.2 1.5 ± 0.4	0.13 ± 0.03 0.13 ± 0.04	0.0020 ± 0.0009 0.0079 ± 0.0029	1.00 ± 0.37 0.95 ± 0.40	200 ± 70 120 ± so	5.4 ± 1.0 12 ± 3	0.16 ± 0.01 0.54 = 0.21	65 ± ″_9 72 ± 38

tamate uptake did not increase with growth irradiance. Because these were mixed cultures (P. piscicida and *Rhodomonas*), it is impossible to determine ab solute uptake rates in either group. However, given the relatively low abundance of **cryptophytes** (ca. 1% of **zoospore** abundance), the contribution of **Rho**domonas to overall uptake rates was likely to be very minor. In fact, applying Stolte and Riegman's (1996) maximum NO; or NH2 uptake rate per cell surface area of 8.5 × 10⁻¹¹ µmol·µm⁻²·h⁻¹, estimates of *Rhodomonas* population (mean area = $45 \mu m^2$) uptake were approximately 0.7 ng at N L^{-1} , h⁻¹, or roughly 0.02% to 0.04% (for NO₃) or 0.5% (for **NH**;) of the measured absolute uptake rates (Table 1). Even in the case of urea, where relatively low uptake rates were measured, theoretical estimates of **Rhodomonas** population uptake (using a maximum /uptake rate of 3 fmol N·cell-1·h-1; Syrett et al. 1986, Price and Harrison 1988) only accounted for 28% or 8% of low or high light estimates, respectively. Therefore, the measured N uptake rates can be accounted for predominantly by P. piscicida uptake.

DISCUSSION

In Stoecker's (1998) review of mixotrophic models, kleptoplastidic protists were considered to acquire nu**trients** predominantly through phagotrophic means, and photosynthesis was thought to be important as a survival mechanism for maintaining respiratory requirements during prey limitation. An inference of this model is that photosynthetically driven cell growth (and therefore associated saprotrophic nu**trient** uptake) is slow relative to that supported by phagotrophy. The potential contribution of photosynthesis to P. piscicida's growth was not addressed in this study, and remains to be determined as a critical test of the model. However, the present findings suggest that saprovoptic uptake can play an important role in P. piscicida's nutrition. The results not only establish that the dinoflagellate can acquire **N-nutrients** directly when kleptoplastidic, but, based on the following arguments, suggest that nutrient **uptake** rates were comparable to those estimated for phytoplankton, and rivaled N uptake through phagotrophy in P. piscicida.

When normalized per cell (based on the entire *P. piscicida* population; i.e. with and without ingested plastids present), uptake rates of NO;, NH⁺, and glutamate (Table 1) were within the range of those reported for phytoplankton (Syrett et al. 1986, Antia et al. 1991, Lomas and Clibert 1999A). Hypotheti-

cally, direct nutrient uptake may have been predominated by, or even confined to, kleptoplastic P. piscicida. Lewitus et al. (1999) presented evidence for greater survival of kleptoplastic cells over apoplastic cells in photosynthetically active P. piscicida cultures. The speculation that a raction of the dinoflagellate population was responsible for the bulk of nutrient uptake would suggest that cellular uptake rates can be potentially higher than those reported here.

Based on previous experiments on *P. piscicida* grazing properties (Glasgow et al. 1998, unpubl. data), comparisons can be made between potential nutrient acquisition via phagotrophy and kleptoplastidy. Ingestion rates by P. piscicida zoospores on Cryptomonas sp. LP757, a cryptophyte resembling Rhodomonas in size and morphology, averaged 1.5 cryptophyte cells per zoospore per day over a 16-h period. Using this estimate for ingestion of cryptophytes, assuming that all of the prey contents were ingested per **encounter** (a false assumption, given the myzocytotic mode of feeding), and using an estimate of cryptophyte N content of 1 pmol-cell-1 (Lewitus and Caron 1990), an N uptake rate can be estimated at 54 fmol N per P. piscicida zoospore per hour. Based on a population density equivalent to those measured in this study, phagotrophic N up take would be approximately 0.8 μ g at N L^{-1} . h^{-1} . Though rough, this estimate suggests that rates of N acquisition by kleptoplastidic P. piscicida (Table 1) may approach or even exceed that obtained through grazing.

The disproportional **distribution** of ambient nutrients in experimental cultures precludes meaningful comparisons of substratespecific uptake rates. It is likely that the relatively high **NO**; uptake rates were related to high ambient **NO**; concentrations (Lomas and Glibert 1999, and references therein). Also, it is possible that the relatively low urea uptake rates were a function of catabolic **enzyme** inhibition (e.g. **urease**) by the relatively high ambient **NH**; concentrations (Flynn and Butler 1986, Antia et al, 1991, Berg et al. 1997). Perhaps a more relevant comparison is the effect of growth irradiance on uptake rates for specific substrates. Only urea (absolute and cellular rates) or NH; (cellular rates) were taken up at greater rates at the higher **irradiance** level. Research is needed to determine whether this effect of growth h-radiance reflects N substrate preference in kleptoplastidic P. piscicida.

From microscopic observations of natural **popu**-

PAGE 07

lations, Glasgow (unpubl. data) has observed chloroplast inclusions in Neuse River, North Carolina 'presumptive Pfiesteria' populations for days after its initial bloom period. In light of the association between prey depletion and kleptoplastidy in laboratory batch cultures of P. piscicida (Lewitus et al. 1999), it is possible that such chloroplast inclusions in field populations are indicative of kleptoplastidy. However, the differentiation between kleptoplastidic and actively grazing Pfiesteria remains problematic, and may depend on the development of specific barkers of photosynthetic activity. In this regard, the use of primulin fluorescence as an indicator (and/or measure) of photosynthetic starch production may be useful for assessing the ecological relevance (occurrence, role, and regulation) of kleptoplastidy in Profiscicia.

roplastidy in *P. piscicida*.

As pointed out, the physiological ecology of klcp toplastidic P. piscicida may bear importantly on the mechanism by which nutrients may regulate their growth. In laboratory experiments (Burkholder et al. 1995, 1998, Burkholder and Glasgow 1997. Glasgow et al. 1998), nutrient (N or P) stimulation of P. piscicida's growth has been demonstrated, either in response to elevated phytoplankton prey supply (the so-called "indirect stimulation" by nutrients) or through direct acquisition of substrates, as shown with T-labeled protein hydrolysate ("direct stimulation"). Based on those laboratoty results and direct correlations between "presumptive. Pfiesteria" abundance, phytoplankton abundance, and nutrient concentrations in natural waters (Burkholder et **[al.** 1997, Burkholder and Glasgow **1997**), a working hypothesis depicting seasonal changes in mechainisms of nutrient stimulation in a temporate estuary **such** as the **Neuse** River is presented in Figure 6, Most of the fish kills involving **P. piscicida** occur in summer. However, a presumably important factor determining the extent and magnitude of toxic acdirect precursors of toxic zoospores. Nontoxic zoospore abundance in the Neuse River has been 'shown to co-vary with chlorophyll during spring phytoplankton bloom periods (Burkholder and Glasigow 1997), with the implication that nutrient regu**lation** of phytoplankton biomass also will indirectly control nontoxic P. piscicida abundance ("indirect /nutrient stimulation"). Also, bloom parameters (magnitude, composition) may affect the proportion of nontoxic zoospores that become kleptoplastidic. We hypothesize that the maintenance (i.e. as **zoospores** rather **than** other cell forms such as cysts or amoebae) and/or growth of these kleptoplastidic populations are dependent on the quantity and quality of available nutrients ("direct stimulation"). In this respect, **P. piscicida's** potential to cause fish Mb could depend on the supply of nutrients avail**able** to support seed populations of nontoxic kleptoplastidic zoospores that "fuel" toxic outbreaks.

Whether or how nutrients stimulate P. pisciada is

THE PROPERTY OF THE PROPERTY O

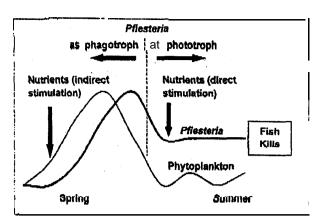


Fig. 6. Model showing hypothesized seasonal changes in mechanisms of nutrient stimulation of *Pfiesteria piscicida*'s growth. In the spring, nutrient acquisition by phagotrophic *P. piscicida* zoospores is derived from phytoplankton prey, and there fore nutrient enrichment may stimulate *P. piscicida*'s growth by increasing the magnitude of phytoplankton blooms ("indirect nutrient stimulation"). After the bloom diipates, nutrients are acquired directly by phototrophic (kleptoplastidic) zoospores, and *P. piscicida*'s growth, or at least maintenance of zoospore populations, may be controlled by the quantity and quality of ambient nutrients ("direct nutrient stimulation"). Both forms of nutrient stimulation would lead to higher abundances of nontoxic zoospores, the direct precursors for toxic zoospores, and therefore increase the potential for toxic outbreaks (e.g. "Fish Kills").

a critical issue in predicting the dinoflagellate's potential impact on estuarine fish populations. The organism is widespread (from Delaware Bay to Mobile Bay, Alabama), but toxic outbreaks have been documented in a relatively narrow range of its latitudinal distribution; that is, in various North Carolina estuaries from 1991-1998, and in the Pocomoke River in summer, 1997 (Burkholder et al. 1995, Lewitus et al. 1995, Burl&older and Glasgow 1997, Maryland Department of Natural Resources 1997, University of Maryland Center for Environmental Science 1997). On comparing regions affected by P. piscicida toxicity with those where P. piscicida is found but not known to cause problems, certain general distinctions in estuarine properties are suggested, including tidal flushing characteristics and fish population dynamics, and also nutrient concentrations (Burkholder and Glasgow 1997). For example, using fish mortality bioassays and scanning electron microscopic confirmation, *P. piscicida was* recently discovered in the pristine North Inlet estuary, South Carolina (Lewitus, Willis, Glasgow, and Burkholder, unpubl. data). In contrast to sites of known P. piscicida toxic events, North Inlet is characterized not only by higher flushing rates, but also by a lack of anthropogenic influence and relatively low inorganic nutrient concentrations (e.g. seasonal maxima in dissolved inorganic nitrogen at some sites rarely exceed 5 µM; Lewitus et al. 1998). Although the North Inlet I? piscicida populations became toxic in fish aquariums, this potential toxicity, to our knowledge, has not been exhibited in a natural habitat. The hypothesized link between high nutrient concentrations and P. piscicida toxic activity suggests that continued coastal eutrophication may lead to an increase in the magnitude and geographical range of P. piscicida toxic events.

Experiments using **P. piscicida zoospores were** conducted at the North Carolina State University Center for Applied Aquatic Ecology. The authors are grateful to Nora Deamer-Melia, Elle Hannon. Erin Haramoto, and Jeff Springer for technical assistance. The research was funded by NSF grants (XX-9403920 and DEB 95-09057, NOAA grant NA90AA-D-SC672, Maryland Department of Natural Resources, Maryland Sea Grant, a CDC graar to South Carolina Department of Health and Environmental Control, and the U.S. ECOHAB Program, sponsored by NOAA, NSF, EPA, NASA, and ONR, grant NA86OP0493. Contribution 1202 of the Belle W. Baruch Institute for Marine Biology and Coastal Research, Contribution 3283 of the University of Maryland Center for Environmental. Science, and ECOHAB Contribution 3.

Antia, N. J., Harrison, P. J. & Oliveira, L. 1991. The role of dissolved organic nitrogen in phytoplankton nunrition, cell biology and ecology. *Phytologia* 30:1-89.

Berg, G. M., Glibert, P. M., Lomas, M. W. & Burford, M. 1997.

Organic nitrogen uptake and growth by the chrysophyte Aureococnus anophagefferens during a brown tide event. Mar. Biol.

227:377-87.

Burke, M. K., Raynal, D. J. & Mitchell, M. J. 1992. Soil nitrogen availability influences seasonal carbon allocation patterns in sugar maple (Act saccharum). Can. J. For. Res. 22:447–56.

Burkholder, J. M. & Glasgow, H. B., Jr. 1997. The ichthyotoxic

dinoflagellate, Pfiesteria piscicida behavior, impacts, and environmental controls. Limnol. Oceanogr. 42:1032-75.

Burkholder, J. M., Glasgow, H. B., Jr. & Hobbs, C. W. 1995. Fish kills linked to a toxic ambush-predator dinoflagellate: distribution and environmental conditions. Mar. Ecol. Prog. Ser. 124;49-61.

Burkholder, J. M., Glasgow, H. B., Jr. & Lewitus, A. J. 1998. The physiological ecology of Pfiesteria pistitin, with general comments on ambush-predator dinoflagellates. In Anderson. D. A., Hallegraeff, G.M. & Cembella, A. D. [Eds.] The Physiological Ecology of Harnful Microalgae. NATO, Paris, pp. 175-91.

Burkholder, J. M., Noga, E. J., Hobbs, C. H. & Glasgow, H. B., Jr. 1992. New 'phantom' dinoflagellate is the causative agent of major estimation fich kills. Nature 359-407 10.

major estuarine fish kills. Nature 358:407-10.

Caron, D. A. 1983. Technique for enumeration of heterotrophic and phototrophic nanoplankton, using epifluorescence microscopy, and comparison with other procedures. Appl. Environ. Microbiol. 46:491-8.

Fields, S. D. & Rhodes, R. G. 1991, Ingestion and retention of Chromonas spp. (Cryptophyceae) by Gymnodinium acidotum (Dinophyceae). J. Phycol. 27:525-9.

Flynn, K. J. & Butler, I. 1986. Nitrogen sources for the growth of

marine microalgae: role of dissolved free amino acids. Mar. Ecol Prog. Ser. 84.981-304

Clasgow, H. B., Jr., Lewitus, A. J. & Burkholder, J. M. 1998. Feeding behavior of the ichthyotoxic estuarine dinoflagellate. Pfiesteria pisciada, on amino acids, al al prey, and fish vs. mammalian erythrocytes. In Reguera, 8.. Blanco, J., Fernandez, M. L. & Wyatt, T. [Eds.] Harnful Microalgae. Proceedings, VIIIth International Conference on Harmful Algal Blooms, Xunta de Calicia and UNESCO, Paris, pp. 3947.

Clibert, P. M. & Capone, D. G. 1999. Mineralization and assimilation in aquatic, sediment and wetland systems. In Knowles, R. & Blackburn, T. H. [Eds.] Nitrogen Isotope Techniques. Academic Press, New York, pp. 245-72.

Guillard, R. R. L. 1975. Culture of phytoplankton for feeding marine invertebrates. In Smith, W. L. & Chanley, M. H.,

[Eds.] Culture of Marine Invertebrate Animals. Plenum, New

Klut, M. E., Stockner, J. & Bisalputra, T. 1939. Further use of

fluorochromes in the cytochemical characterization of phytoplankton. Histochem. J. 21:645-50.

Larsen, J. 1992. Endocytobiotic consortia with dinoflagellate hosts. In Reisser, W. [Ed.] Algae and Symbioses: Plants, Animals, Fungi, Viruses, Interactions Expland. Biopress Limited. Bristol. DD. 427~42.

Laval-Peuto, M. 1992. Plastidic protozoa. In Reisser, W. [Ed.] Algae and Symbioses: Plants, Animals, Fungi, Viruses, Interactions Explored. Biopress Limited. Bristol. pp. 471-99.

Lewitus, A. J. & Caron, D. A. 1990. Relative effects of nitrogen or phosphorus depletion and light intensity on the pigmentation, chemical composition, and volume of Pyrenomonas sa-

lina (Cryptophyceae). Mar. Ecol. Prog. Ser. 61:171-81.
Lewitus, A. J., Glasgow, H. B., Jr. & Burkholder, J. M. 1999. Kleptoplastidy in the toxic dinoflagellate. Priesteria piscicida (Di-

nophyceae). J. Phycol. 35:303-12.

Lewitus, A. J... Jesien, R. V. Kana. T. M.. Burkholder, J. M., Clasgow, H. B., Jr. & May, E. 1995. Discovery of the "phantom" dinoflagellate in Chesaeake Bay. Estuaries 18:373-8.

Lewitus, A. J., Koepfler, E. T. & Morris, J. T. 1998. Seasonal variation in the regulation of phytoplantam, by nitrogen and

iation in the regulation of phytoplankton by nitrogen and grazing in a salt-marsh estuary. Limnol Oceanogr. 43:636-46.

Lindroth, P. & Mopper. K. 1979. High performance liquid chromatographic determination of subpicomole amounts of amino acids by precolumn fluorescence derivatization with ophthaldialdehyde. Analyt. Chem. 51:1667-74.

Lomas, M. W. Sc Clibert, P. M. 1999. Interactions between NH, and NO, uptake and assimilation: comparison of diatoms and dinoflagellates at several growth temperatures. Mar. Biol. (193:541-56).

-1999 Temperature regulation of nitrate uptake: A novel hypothesis about nitrate uptake and reduction in cool-water diatoms. Limnol. Oceanogr (44:556-72).

Maryland Department of Natural Resources. 1997. Water quality, habital and biological conditions of river systems affected by Pfiesteria or Pfiesteria-like organisms on the Lower Eastern Shore Of Maryland. 1997 Summary, Tidewater Ecosystem Assessment Division Report.

Parsons, T. R., Maita, Y. & Lalli, C. M. 1984. A manual of chemical and biological methods for seawater analysis. Pergamon Press, Ox-

Porter. K. G. & Feig, Y. S. 1980. The use of DAPI for identifying and counting aquatic microflora. Limnol. Oceanogr. 25:943-8.

Price. N. M. & Harrison, P. J. 1988. Uptake of urea C and urea
N by the coastal marine diatom Thalassiosira pseudonana. Lim-

nol. Oceanogr. 33:528-37.
Revilla, M. A.: Tolivia, D. & Tarrago, J. F. 1986. A new and permanent staining method for starch granules using fluorescence microscopy. Slain Technol. 61:151-4.

Santore. U. J. 1985. A cytological survey of the genus Criptomonus (Cryptophyceael with comments on its taxonomy. Arch. Pro-tistent. 190:1-52.

Schnepf, E. & Elbráchter, M. 1992. Nutritional strategies in dinoflagellates. A review with emphasis on cell biological aspects, Eur. J. Protistol. 28:3-24.

Schnepf, E., Winter, S. & Mollenhauer, D. 1989. Cymnodinium neruginosum (Dinophyta): a blue-green dinoflagellate with a vestigial, anucleate, cryptophycean endosymbiont. Plant Syst. Evol. 164:75-91.

Sieracki, M. E., Haas, L., Caron, D. A. & Lessard, E. J. 1987. The effect of fixation on particle retention by microflagellates: underestimation of grazing rates. Mar. Ecol. Prog. Ser. S&231-

Skovgaard, A. 1998. Role of chloroplast retention in a marine

dinoflagellate. Aquat. Microb. Ecol. 15:293-301.
Solorzano, L. 1969. Determination of ammonia in natural waters by the phenohypochlorite method. Limnol. Oceanugr, 14:16-

Steidinger, K. A., Burkholder, J. M., Glasgow H. B., Jr., Hobbs. C. W., Garrett, J. K., Truby, E. W., Noga, E. J. & Smith, S. A. 1996. Pfiesteria piscicida gen. Et sp. now (Pfiesteriaceae fam.

1437

NITROGEN UPTAKE BY PITESTERIA

nov.), a new toxic dinoflagellate with a complex life cycle and behavior. J. Physol. 32:157-64.

Sterling, C. 1964. Starch-primulin fluorescence, Protoplasma 59: 180-92.

Stoecker, D. K. 1998. Conceptual models of mixotrophy in planktonic protists and some ecological and evolutionary implications. Eur. J. Protistol. 34:28 I-290.

tive competition of marine phytoplankton for fluctuating nitrate and ammonium. J. Physol. 32732-K).

Syrett, P. J., Flynn, K. J., Molloy, C. J., Dixon. C. K., Peplinska, A. M. & Gresswell, R. C. 1986. Effects of nitrogen deprivation on rates of uptake of nitrogenous compounds by the diatom, Phaeodactylum incornutum Bohlin. New Phytol. 102:39-44.

Throndsen, J. 1978. Preservation and Storage. In Sournia. A, (Ed. 1. Dhylothambton Manual LINESCO Paris no. 69-74.